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Changes in the activities of glutathione-S-transferases, glutathione reductase and catalase after exposure to different concentrations of cadmium in *Australoheros facetus* (Cichlidae, Pisces)

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Abstract

The goal of this work was to evaluate the response of the glutathione-S-transferases (GST), glutathione reductase (GR) and catalase (CAT) enzymes in different tissues of the fish *Australoheros facetus* exposed cadmium (Cd). In liver, a significant decrease of GST activity was observed at 0.031 mg L⁻¹ (p<0.05), while, at higher concentrations, there were non-significant changes compared to controls. Neither the GR nor the CAT activity was different from control. Gills was the most sensitive organ, showing increased GST activity at 0.31 and 1.53 mg L⁻¹ and GR at 3.06 mg L⁻¹; while CAT activity was inhibited at 0.31, 3.06 and 6.13 mg L⁻¹ (p<0.05). The brain GST activity was increased at 1.53 mg L⁻¹ but inhibited at 3.06 and 6.13 mg L⁻¹, meanwhile GR and CAT activities were inhibited at 1.2, 3.06 and 6.13 mg L⁻¹ (p<0.05). The induction of defense enzymes as well as the inhibition of the antioxidant enzyme catalase, even at environmentally relevant concentrations used in this work (e.g. 0.31 mg L⁻¹), denotes the toxic effect that cadmium exerts over the tissues of *A. facetus* and alerts over the need of doing biomonitoring in areas potentially polluted with metals.

Keywords: biomarkers, cadmium, catalase, glutathione reductase, glutathione-S-transferases.

Cambios en las actividades de glutathione-S-transferasa, glutathione reductasa y catalasa en *Australoheros facetus* (Cichlidae, Pisces) expuesto a diferentes concentraciones de cadmio

Resumen

El objetivo de este trabajo fue evaluar la respuesta de las enzimas glutathione-S-transferasa (GST), glutathione reductasa (GR) y catalasa (CAT) en diferentes tejidos del pez *Australoheros facetus* expuesto a cadmio (Cd). En hígado, se observó una disminución de la actividad de GST a 0,031 mg L⁻¹ (p<0,05), mientras que a concentraciones mas elevadas no se hallaron diferencias respecto al control. La branquia fue el órgano mas sensible, presentando un incremento de actividad de GST a 0,31 y 1,53 mg L⁻¹ Cd y de GR a 3,06 mg L⁻¹ Cd; mientras que la actividad de CAT resultó inhibida a 0,31, 3,1 y 6,13 mg L⁻¹ Cd (p<0,05). En el cerebro la actividad de GST se incrementó a 1,53 mg L⁻¹ Cd y se inhibió a 3,1 y 6,13 mg L⁻¹ Cd, mientras que las actividades de GR y CAT se inhibieron a 1,53, 3,1 y 6,13 mg L⁻¹ Cd (p<0,05). La inducción de los sistemas de defensa así como la inhibición del enzima antioxidante catalasa, aún a concentraciones de relevancia ambiental (ejemplo 0,31 mg L⁻¹), demuestra el efecto tóxico que ejerce el Cd en los tejidos de *A. facetus* y alerta sobre la necesidad de realizar biomonitoreos en áreas potencialmente contaminadas con metales pesados.

Palabras clave: biomarcadores, cadmio, catalasa, glutathione reductasa, glutathione-S-transferases.

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INTRODUCTION

Metals can cause an increase in reactive oxygen species (ROS) leading to oxidative damage. Cadmium (Cd) binds to complex III in the electron transport chain promoting the production of ROS (Boelsterli, 2007). Under oxidative stress, antioxidant defenses are induced to protect animal organisms from free radicals as the first line of defense against Cd, before the induction of any other detoxification mechanism (Basha & Rani, 2003). In the aquatic environment fishes are exposed to different classes of pollutants being the metals one of the major group of aquatic contaminants. In fact, all water bodies are polluted by metals, elements as mercury (Hg), cadmium (Cd), copper (Cu), and zinc (Zn) are considered the most dangerous in the ecotoxicological aspect (Golovanova, 2008). Although in general the environmental levels of Cd do not exceed guideline values ($<0.25 \text{ ug L}^{-1}$, SRHN, 2003), aquatic organisms are able to accumulate this chemical up to concentrations that are tens and even thousands of times higher than their concentrations in the environment (Golovanova, 2008). Therefore, sublethal concentrations in water can lead finally to toxic effects in biota. In particular, values of cadmium as high as 0.7- 1.7 mg/L have been reported in surface water in Argentina (Salibián, 2006) alerting about its potential effects in the freshwater biota.

Australoheros facetus is a common freshwater species distributed in Argentina, Brazil, Uruguay and Paraguay (Rican & Kullander, 2006). It is easy to maintain, grow and reproduce in the laboratory (Bulus Rossini & Ronco, 2004). It is sensitive to metals and to Cd in particular, showing a LC_{50} of 0.091 mg L^{-1} in 15-days old (post-hatch) organisms, dying within 96 h. When comparing the relative sensitivity of this species, the following trend, from the most to the least toxic, can be seen: $\text{Cd(II)} > \text{Hg(II)} > \text{Cu(II)} > \text{Zn(II)} > \text{Cr(VI)}$ (Bulus Rossini & Ronco, 2004). Our work is the first report concerning sublethal effects of Cd in adult organisms from the same species.

Taking into account that activities of enzymes can vary greatly among different tissues within the same organism (Huggett *et al.*, 1992) the goal of the present work was to evaluate changes in the activities of glutathione-S-transferases (GST), glutathione reductase (GR) and catalase (CAT) enzymes in liver, gill and brain tissues of the freshwater fish species *Australoheros facetus* exposed to environmentally relevant and higher concentrations of Cd.

MATERIALS AND METHODS

Fish exposure

Fish were collected in freshwater bodies around Mar del Plata city (Province of Buenos Aires, Argentina, $37^{\circ} 53'$ South, $57^{\circ} 59'$ West) and acclimatized to laboratory conditions prior to the experiments for one month in 140-L tanks. The experimental room was illuminated with fluorescent lamps with 12 h light periods. The tap water used for the experiments had a temperature of $20 \pm 1^{\circ}\text{C}$, pH values of 8.2 ± 0.2 , a total hardness of $374 \text{ mg L}^{-1} \text{ CaCO}_3$ and alkalinity of $160 \text{ mg L}^{-1} \text{ CaCO}_3$. A stock solution of 1 g L^{-1} of cadmium chloride (CdCl_2) (Sigma) was prepared in distilled water and aliquots were taken to prepare different exposure solutions in a final volume of 30 L. The experiment was conducted in glass aquaria, each containing 6 fishes ($n=6$ per treatment) in 30 L of tap water (controls) and in contaminated test water. Fish were exposed during 24 h to 0,005, 0.05, 0.5, 2.5, 5 and 10 mg L^{-1} of CdCl_2 being the final concentrations of Cd: 0, 0.0031, 0.031, 0.31, 1.53, 3.06 and 6.13 mg L^{-1} , respectively. At the end of the exposure period, all fishes were dissected using a fresh razor blade. Livers, gills and brains were sampled, weighed, immediately frozen using liquid nitrogen, and stored at -80°C until analysis. The morphometric characteristics are shown in Table 1.

Enzyme extraction

The extraction of cytosolic enzymes was done according to the method described by Wiegand *et al.* (2000) without the purification step. Briefly, 50 mg of each tissue material were used. Frozen tissues were homogenized using a Potter glass homogenizer (packed in ice) with 5 mL sodium-phosphate buffer (0.1 M, pH 6.5) containing 20% glycerol, 14 mM 1,4- dithioerythritol (DTE), and 1 mM EDTA. Cell debris was removed by centrifugation at $10,000 \text{ g}$ for 10 min. The supernatant was centrifuged at $100,000 \text{ g}$ for 60 min to separate the membrane- associated fraction from the soluble fraction. The cytosolic fraction was immediately frozen using liquid nitrogen, and stored at -80°C until measurement

Enzymatic activities were determined by spectrophotometry. The activity of the soluble (cytosolic) glutathione S-transferase (GST) was determined using 1-chloro-2, 4-dinitrobenzene (CDNB) as substrate, according to Habig *et al.* (1974). Glutathione reductase activity (GR) was assayed according to Tanaka *et al.* (1994) and catalase activity (CAT) according to Claiborne (1985). The enzymatic activities were estimated in

Table 1: Morphometric characteristics of *Australoheros facetus* after exposure to cadmium. Data are mean \pm standard deviation.

Treatment	Total Length (cm)	Total Weight (g)	Condition Factor (K)	Hepatic Index
Control	9.7 ± 0.8	22.6 ± 6.3	2.39 ± 0.1	4.6 ± 1.0
0.0031 mg L^{-1}	9.7 ± 0.5	22.4 ± 3.5	2.40 ± 0.1	4.3 ± 0.6
0.031 mg L^{-1}	9.9 ± 0.3	24.4 ± 2.7	2.47 ± 0.1	5.0 ± 0.8
0.31 mg L^{-1}	9.8 ± 0.9	24.1 ± 6.5	2.45 ± 0.1	4.4 ± 0.8
1.53 mg L^{-1}	10.1 ± 1.0	26.3 ± 7.6	2.48 ± 0.1	4.5 ± 1.5
3.06 mg L^{-1}	10.0 ± 1.0	23.8 ± 7.2	2.28 ± 0.2	4.2 ± 0.9
6.13 mg L^{-1}	9.7 ± 0.9	24.0 ± 7.0	2.53 ± 0.3	4.1 ± 1.0

terms of the protein content for each sample (Bradford 1976), using a bovine serum albumin solution as standard. Enzymatic activities are reported in nano katal per milligram of protein (nkat. mg⁻¹ prot), where 1 katal correspond to the conversion of 1 mol of substrate per second. Each enzymatic assay was carried out in duplicate. All reactive were purchased from Sigma.

Statistics

Differences between treatments were tested by non-parametric Kruskal-Wallis test and *a posteriori* by the Mann-Whitney U-test (Zar, 1999).

RESULTS

Liver

A significant decrease of GST activity was observed at 0.031 mg L⁻¹ Cd (p<0.05), while, at higher concentrations, there were non-significant changes compared to controls. Neither the GR nor the CAT activity was different from controls (p>0.05) (Fig. 1).

Gills

Increase of GST activity was observed at 0.31 and 1.53 mg L⁻¹ Cd concentrations (Fig. 2). Glutathione reductase activity was significantly increased at 3.1 mg L⁻¹ while CAT activity was significantly decreased at 0.31, 3.1 and 6.13 mg L⁻¹ (p<0.05) (Fig. 2).

Brain

Increase of GST activity at 1.53 mg L⁻¹ Cd was observed (p<0.05) while at 3.1 and 6.13 mg L⁻¹ the activity was lower

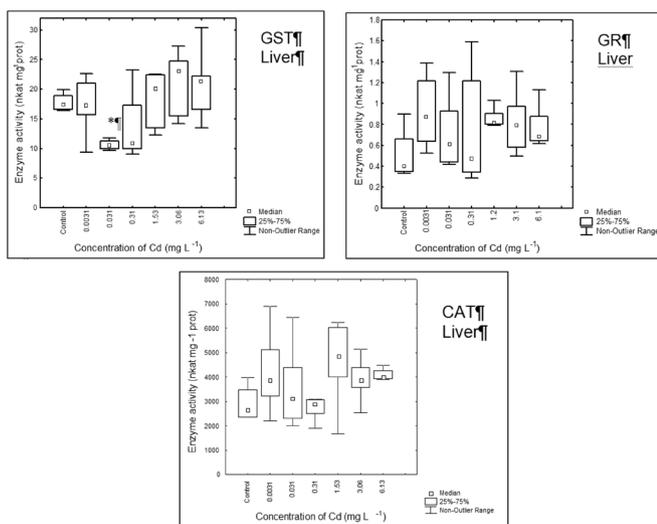


Figure 1: Concentration-response curve in liver of *Australoheros facetus* after exposure to cadmium. Data are median \pm standard error. Exposure time: 24 h. *: significantly different to control (p<0.05). GST: glutathione-S-transferase, GR: glutathione reductase, CAT: catalase.

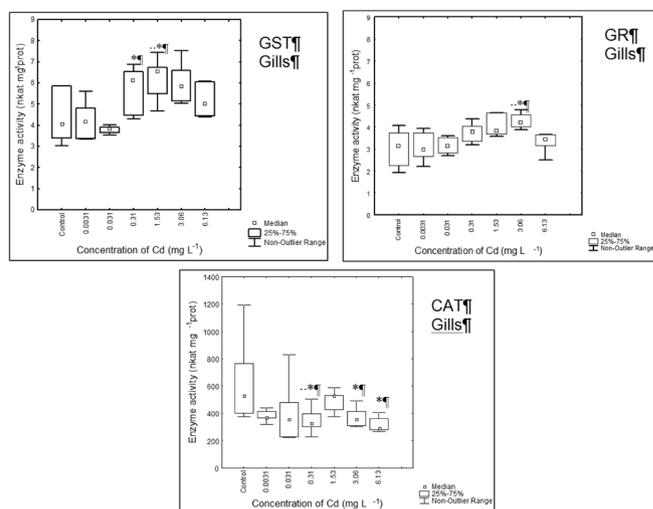


Figure 2: Concentration-response curve in gills of *Australoheros facetus* after exposure to cadmium. Data are median \pm standard error. Exposure time: 24 h. *: significantly different to control (p<0.05). GST: glutathione-S-transferase, GR: glutathione reductase, CAT: catalase.

than in control (p<0.05) (Fig. 3). At concentrations of 0.31, 1.53, 3.1 and 6.13 mg L⁻¹ Cd, a 1.4- 2.7-fold significant decrease of GR activity was detected (p<0.05) (Fig. 3). A significant 2.1- 2.6- fold decrease of CAT activity at Cd concentrations of 1.53, 3.1 and 6.13 mg L⁻¹ was observed (p<0.05) (Fig. 3).

DISCUSSION

It has been previously shown that Cd exposure alters GSH/GSSG ratio in fish species such as rainbow trout (Lange *et al.*, 2002). Glutathione-S-transferases constitute a complex family of proteins that play roles in both normal cellular metabolism and in the detoxification of a wide variety of xenobiotic compounds. They are predominantly cytosolic defence systems responsible for protecting cellular components against various toxic effects and oxidative stress (Sen & Semiz, 2007). It has also been demonstrated that the role of GST in oxidative stress is of conjugating endogenously-produced electrophiles such as membrane lipid peroxides in animals (Halliwell & Gutteridge, 1999) as well as in plants (Cummins *et al.*, 1999). According to our present results, GST and GR could play a role in oxidative stress tolerance and could act protecting *A. facetus* against damages induced by Cd, especially when other defence systems are not efficient, like CAT that was inhibited even with low Cd concentrations (e.g. 0.31 mg L⁻¹ in gills). Cadmium induced enzymatic defences and when the enzyme activities are overwhelmed inhibited damage could occur. Although damage in *A. facetus* was not studied in this work, it represents a potential effect because oxidative damage measured as increased malondialdehyde (MDA) content as well as DNA fragmentation (Jia *et al.*, 2010) and micronucleous (Ossana *et al.*, 2009) have been reported at 0.5 mg L⁻¹ Cd in other fish species.

Brain is particularly susceptible to oxidative damages. In spite of the high rate production of ROS, from the high rate

oxidative metabolism and abundance of polyunsaturated fatty acids in cell membrane, brain has a relatively low antioxidant defence system (Matés, 2000; Verstraeten *et al.*, 2008). Although it has been reported that brain is not the main target organ of Cd in rats or that this metal it is not able to pass through the blood-brain barrier (Carrasco Trancoso, 2000), brain represent a sensitive target organ to Cd toxicity in *A. facetus*. This fact is clearly evidenced by the inhibition of the three enzymatic systems analyzed in the organ. Catalase has been shown to be either induced or inhibited by metals depending on the dose, the species and the route of exposure (Atli *et al.*, 2006). The response of CAT observed in gills and brain could be due to the increased flux of superoxide radicals and hydrogen peroxide, which had been reported to inhibit CAT activity in other freshwater fishes possibly due to the direct binding of metal ions to -SH groups of the enzyme molecule (Pandey *et al.*, 2001; Atli *et al.*, 2006).

For many xenobiotics, the underlying mechanisms that govern their organotrophy or make a particular tissue prone to succumb to the resulting toxic effects are not known (Boelsterli, 2007). Fish gills constitute a direct exposure route for environmental water contaminants and it is the first organ affected when fishes are exposed to metals (Atli *et al.*, 2006). In this study gills were more sensitive than liver. Also, brain has been proposed as more sensitive than liver and one of the most important target organs in fishes (Song *et al.*, 2006). Because of that, both organs could be considered in future biomarkers evaluations. Our results extend the current knowledge by suggesting the participation of GST and GR as defence mechanisms against Cd in *A. facetus* since increased enzyme activities were observed in response to Cd exposure. On the other hand, the inhibition of CAT observed in gills and brain showed the oxidative stress caused by this metal.

Overall, the induction of defense enzymes as well as the inhibition of the antioxidant enzyme catalase, even at

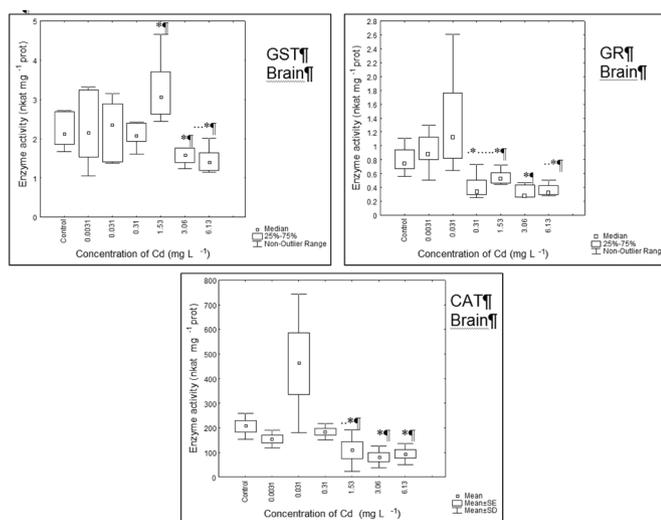


Figure 3: Concentration-response curve in brain of *Australoheros facetus* after exposure to cadmium. Data are median \pm standard error. Exposure time: 24 h. *: significantly different to control ($p < 0.05$). GST: glutathione-S-transferase, GR: glutathione reductase, CAT: catalase.

environmentally relevant concentrations used in this work, denotes the toxic effect that cadmium exerts over the tissues of *A. facetus* and alerts over the need of doing biomonitoring in areas potentially polluted with metals.

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